

Ecography

ECOG-04374

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Supplementary material

Human overexploitation and extinction risk correlates of Chinese snakes

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17 **Appendix 1. Properties of the datasets used, hypotheses and justification**

18 **Table A1.** Extinction risk (based on China Biodiversity Red List), intrinsic traits and extrinsic factors of Chinese snakes. Abbreviations: China,
 19 China Biodiversity Red List; RangeC, species assessed under range-based criteria; IUCN, IUCN Red List; BL, Body length; BR, Body ratio;
 20 AP, Activity period; MH, Microhabitat; RM, Reproductive mode; HS, Habitat specificity; RS, Range size; MT, Mean temperature; MP, Mean
 21 precipitation. See Minimum elevation (ME) and Exploitation index (EI) in Table S3.

| Species ^a | Family ^a | China ^b | RangeC ^c | IUCN ^b | BL (mm) | BR ^d | Toxicity ^e | AP ^f | MH ^g | RM ^h | HS ⁱ | RS | MT | MP |
|---------------------------------|---------------------|--------------------|---------------------|-------------------|---------|-----------------|-----------------------|-----------------|-----------------|-----------------|-----------------|----------|---------|-----------|
| <i>Achalinus ater</i> | Xenodermatidae | LC | 0 | LC | 327.5 | 1.0216 | 0 | 1 | 3 | 0 | 2 | 6.5275 | 17.0857 | 1275.5387 |
| <i>Achalinus formosanus</i> | Xenodermatidae | LC | 0 | LC | 720 | 1.0000 | 0 | 1 | 3 | 0 | 1 | 0.0543 | 20.5532 | 1098 |
| <i>Achalinus hainanus</i> | Xenodermatidae | VU | 1 | VU | 310 | 1.0000 | 0 | 1 | 3 | 0 | 2 | 0.0341 | 24 | 1500 |
| <i>Achalinus jinggangensis</i> | Xenodermatidae | VU | 1 | CR | 407.5 | 0.7717 | 0 | 1 | 3 | 0 | 1 | 0.0278 | 17.5 | 1836 |
| <i>Achalinus meiguensis</i> | Xenodermatidae | LC | 0 | LC | 520 | 0.8739 | 0 | 1 | 3 | 0 | 3 | 12.0946 | 11.3599 | 950.9581 |
| <i>Achalinus niger</i> | Xenodermatidae | LC | 0 | LC | 588 | 1.0381 | 0 | 1 | 3 | 0 | 2 | 2.0545 | 20.4199 | 1102.2115 |
| <i>Achalinus rufescens</i> | Xenodermatidae | LC | 0 | LC | 400 | 0.9048 | 0 | 1 | 3 | 0 | 1 | 57.0637 | 19.8764 | 1599.2118 |
| <i>Achalinus spinalis</i> | Xenodermatidae | LC | 0 | LC | 515 | 0.8393 | 0 | 1 | 3 | 0 | 2 | 124.4304 | 15.9276 | 1269.8047 |
| <i>Ahaetulla prasina</i> | Colubridae | LC | 0 | LC | 1133 | 1.0071 | 1 | 3 | 5 | 1 | 3 | 36.0525 | 20.1028 | 1456.8462 |
| <i>Amphiesma stolatum</i> | Colubridae | LC | 0 | NA | 616 | 0.8667 | 0 | 3 | 2 | 0 | 4 | 197.0599 | 17.3927 | 1309.2899 |
| <i>Amphiesmoides ornaticeps</i> | Colubridae | VU | 1 | LC | 846.5 | 1.0155 | 0 | 2 | 2 | 1 | 1 | 13.3319 | 21.0513 | 1614.5862 |
| <i>Archelaphe bella</i> | Colubridae | VU | 1 | LC | 1009 | 1.0697 | 0 | 3 | 3 | 0 | 1 | 51.4 | 16.8772 | 1182.0268 |
| <i>Atretium yunnanensis</i> | Colubridae | LC | 0 | LC | 815 | 0.7285 | 0 | 3 | 2 | 0 | 2 | 4.2593 | 11.2851 | 1280.4043 |
| <i>Azemiope feae</i> | Viperidae | VU | 0 | LC | 666 | 1.0151 | 1 | 2 | 3 | 0 | 4 | 269.679 | 10.7868 | 710.011 |
| <i>Boiga cyanea</i> | Colubridae | VU | 1 | NA | 1290 | 0.8169 | 1 | 1 | 4 | 0 | 1 | 39 | 16.3175 | 1039.7943 |
| <i>Boiga guangxiensis</i> | Colubridae | VU | 1 | LC | 1685 | 0.9824 | 1 | 1 | 5 | 0 | 1 | 1.582 | 20.9128 | 1554.3605 |
| <i>Boiga kraepelini</i> | Colubridae | LC | 0 | LC | 1404 | 1.0014 | 1 | 1 | 5 | 0 | 2 | 159.4006 | 17.4836 | 1369.1565 |
| <i>Boiga multomaculata</i> | Colubridae | LC | 0 | NA | 790 | 0.8545 | 1 | 1 | 5 | 0 | 2 | 162.7306 | 17.8597 | 1382.0514 |

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|----------------------------------|------------|----|---|----|--------|--------|---|---|---|---|---|----------|---------|-----------|
| <i>Bungarus fasciatus</i> | Elapidae | EN | 0 | LC | 1486 | 0.945 | 1 | 1 | 3 | 0 | 3 | 5.2808 | 7.7328 | 660.5455 |
| <i>Bungarus multicinctus</i> | Elapidae | EN | 0 | LC | 1492.5 | 1.0874 | 1 | 1 | 3 | 0 | 3 | 147.8303 | 17.8776 | 1369.4992 |
| <i>Calamaria pavementata</i> | Colubridae | LC | 0 | LC | 314.5 | 2.3105 | 0 | 1 | 3 | 0 | 2 | 108.4886 | 17.9549 | 1333.765 |
| <i>Calamaria septentrionalis</i> | Colubridae | LC | 0 | LC | 353.5 | 0.846 | 0 | 1 | 3 | 0 | 2 | 110.6583 | 18.2639 | 1484.7749 |
| <i>Calamaria yunnanensis</i> | Colubridae | VU | 1 | EN | 353.5 | 0.6557 | 0 | 1 | 3 | 0 | 2 | 0.3562 | 18 | 909.3 |
| <i>Chrysopelea ornata</i> | Colubridae | VU | 1 | NA | 807.5 | 1.1824 | 1 | 3 | 5 | 0 | 2 | 55.0506 | 17.3346 | 1225.138 |
| <i>Coelognathus radiatus</i> | Colubridae | EN | 0 | LC | 1556 | 0.9329 | 0 | 4 | 3 | 0 | 4 | 110.7673 | 17.9901 | 1322.9935 |
| <i>Cyclophiops doriae</i> | Colubridae | VU | 1 | NA | 925 | 1.0000 | 0 | 3 | 4 | 0 | 3 | 0.4428 | 19.95 | 2024.6 |
| <i>Cyclophiops major</i> | Colubridae | LC | 0 | LC | 1314 | 0.9879 | 0 | 3 | 4 | 0 | 4 | 185.4334 | 20.3469 | 1115.4491 |
| <i>Cyclophiops multicinctus</i> | Colubridae | NT | 0 | LC | 1251 | 1.0177 | 0 | 3 | 4 | 0 | 1 | 15.1129 | 20.5347 | 1445.4099 |
| <i>Daboia siamensis</i> | Viperidae | EN | 1 | LC | 1009 | 1.0000 | 1 | 4 | 3 | 1 | 5 | 117.8393 | 18.2166 | 1364.5699 |
| <i>Deinagkistrodon acutus</i> | Viperidae | EN | 0 | NA | 1286.5 | 1.0784 | 1 | 4 | 3 | 0 | 6 | 203.536 | 17.3576 | 1315.5441 |
| <i>Dendrelaphis hollinrakei</i> | Colubridae | NT | 0 | DD | 1200 | 1.0000 | 0 | 3 | 5 | 0 | 2 | 0.1106 | 22 | 1912.8333 |
| <i>Dendrelaphis pictus</i> | Colubridae | LC | 0 | NA | 1334 | 0.8618 | 0 | 3 | 5 | 0 | 2 | 84.2906 | 18.6357 | 1343.3129 |
| <i>Elaphe anomala</i> | Colubridae | VU | 0 | NA | 2135 | 0.9409 | 0 | 4 | 3 | 0 | 2 | 322.1546 | 8.8474 | 565.9058 |
| <i>Elaphe bimaculata</i> | Colubridae | LC | 0 | LC | 846.5 | 0.8402 | 0 | 4 | 3 | 0 | 3 | 175.7358 | 12.9147 | 881.8269 |
| <i>Elaphe carinata</i> | Colubridae | EN | 0 | NA | 2147.5 | 1.0452 | 0 | 4 | 3 | 0 | 2 | 380.2246 | 14.1715 | 904.0103 |
| <i>Elaphe davidi</i> | Colubridae | VU | 0 | NA | 1180 | 1.4842 | 0 | 4 | 3 | 0 | 2 | 253.9446 | 5.8604 | 413.8952 |
| <i>Elaphe dione</i> | Colubridae | LC | 0 | LC | 1007.5 | 0.9639 | 0 | 4 | 3 | 0 | 6 | 400.9079 | 6.1971 | 446.1386 |
| <i>Elaphe schrenckii</i> | Colubridae | VU | 0 | NA | 1587 | 1.0165 | 0 | 4 | 3 | 0 | 4 | 80.84 | 3.5157 | 539.7987 |
| <i>Elaphe zoigeensis</i> | Colubridae | LC | 0 | LC | 880 | 1.0000 | 0 | 3 | 3 | 0 | 1 | 1.0437 | 1.1 | 648.5 |
| <i>Emydocephalus ijimae</i> | Elapidae | LC | 0 | LC | 739.5 | 0.7079 | 1 | 4 | 1 | 1 | 1 | 0.1866 | 20.833 | 1520.7273 |
| <i>Eryx miliaris</i> | Boidae | VU | 0 | NA | 778.5 | 0.5937 | 0 | 2 | 3 | 1 | 2 | 336.31 | 6.4199 | 185.4227 |
| <i>Eryx tataricus</i> | Boidae | VU | 0 | NA | 409 | 0.9569 | 0 | 2 | 3 | 1 | 2 | 336.31 | 6.4199 | 185.4227 |
| <i>Euprepiophis mandarinus</i> | Colubridae | VU | 0 | LC | 1332.5 | 1.1492 | 0 | 4 | 3 | 0 | 3 | 252.81 | 20.0313 | 1113.4769 |
| <i>Euprepiophis perlacea</i> | Colubridae | EN | 1 | EN | 1197 | 0.9244 | 0 | 4 | 3 | 0 | 1 | 4.0435 | 10.9729 | 1097.7585 |
| <i>Gloydus brevicaudus</i> | Viperidae | NT | 0 | NA | 624 | 1.0032 | 1 | 2 | 3 | 1 | 2 | 356.6846 | 12.9808 | 798.6924 |
| <i>Gloydus intermedius</i> | Viperidae | NT | 0 | NA | 595 | 0.9833 | 1 | 3 | 3 | 1 | 2 | 444.66 | 6.0848 | 236.1956 |

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|--------------------------------|------------|----|---|----|--------|--------|---|---|---|---|---|----------|---------|-----------|
| <i>Gloydus lijianlii</i> | Viperidae | VU | 1 | NA | 687 | 1.0000 | 1 | 3 | 5 | 1 | 2 | 0 | 12.6 | 672.5 |
| <i>Gloydus liupanensis</i> | Viperidae | NT | 0 | NA | 549 | 1.1115 | 1 | 3 | 3 | 1 | 3 | 52.01 | 7.1354 | 262.2981 |
| <i>Gloydus monticola</i> | Viperidae | NT | 0 | DD | 433.5 | 0.7202 | 1 | 2 | 3 | 1 | 2 | 39 | 16.3175 | 1039.7943 |
| <i>Gloydus qinlingensis</i> | Viperidae | NT | 0 | NA | 531 | 0.9594 | 1 | 3 | 3 | 1 | 3 | 227.3923 | 6.653 | 374.9583 |
| <i>Gloydus shedaoensis</i> | Viperidae | EN | 1 | VU | 747.5 | 1.0479 | 1 | 3 | 5 | 1 | 2 | 0.0091 | 10 | 575 |
| <i>Gloydus strauchi</i> | Viperidae | NT | 0 | NA | 496.5 | 0.9133 | 1 | 3 | 3 | 1 | 4 | 232.3023 | 6.7891 | 385.1617 |
| <i>Gloydus ussuriensis</i> | Viperidae | NT | 0 | NA | 647 | 1.054 | 1 | 2 | 3 | 1 | 4 | 80.84 | 3.5157 | 539.7987 |
| <i>Gonyosoma boulengeri</i> | Colubridae | VU | 1 | NA | 1206.5 | 1.1354 | 0 | 3 | 4 | 0 | 1 | 4.4002 | 22.2154 | 1706.7078 |
| <i>Gonyosoma frenatus</i> | Colubridae | LC | 0 | NA | 1342.5 | 1.219 | 0 | 3 | 5 | 0 | 4 | 182.6967 | 15.5845 | 1007.2159 |
| <i>Gonyosoma prasinus</i> | Colubridae | VU | 1 | NA | 908.5 | 1.0144 | 0 | 4 | 3 | 0 | 1 | 35.879 | 17.2714 | 1144.6999 |
| <i>Hebius atemporale</i> | Colubridae | NT | 0 | DD | 444 | 1.1926 | 0 | 3 | 3 | 0 | 2 | 33.5054 | 19.8577 | 1465.8178 |
| <i>Hebius bitaeniatum</i> | Colubridae | NT | 0 | LC | 634 | 0.9875 | 0 | 1 | 3 | 0 | 1 | 15.8145 | 18.0489 | 1140.2754 |
| <i>Hebius boulengeri</i> | Colubridae | LC | 0 | LC | 584.5 | 1.0155 | 0 | 1 | 2 | 0 | 3 | 47.1533 | 20.2865 | 1575.7067 |
| <i>Hebius craspedogaster</i> | Colubridae | LC | 0 | LC | 466.5 | 0.8697 | 0 | 3 | 2 | 0 | 4 | 175.0492 | 16.7594 | 1298.651 |
| <i>Hebius johannis</i> | Colubridae | LC | 0 | NA | 768.5 | 0.9655 | 0 | 1 | 2 | 0 | 2 | 105.2167 | 13.5089 | 611.7649 |
| <i>Hebius metusium</i> | Colubridae | NT | 0 | EN | 757.5 | 0.7119 | 0 | 3 | 2 | 0 | 2 | 0.2062 | 12.9013 | 1325.7 |
| <i>Hebius miyajimae</i> | Colubridae | EN | 1 | VU | 513 | 0.9731 | 0 | 4 | 2 | 0 | 1 | 1.0153 | 19.5647 | 1145.537 |
| <i>Hebius modestum</i> | Colubridae | LC | 0 | LC | 675 | 0.8 | 0 | 1 | 3 | 0 | 2 | 0.1135 | 9.7107 | 858.4286 |
| <i>Hebius octolineatum</i> | Colubridae | LC | 0 | LC | 834.5 | 0.9751 | 0 | 1 | 2 | 0 | 5 | 39.5003 | 14.3431 | 944.2623 |
| <i>Hebius optatum</i> | Colubridae | LC | 0 | NA | 700 | 0.9718 | 0 | 3 | 2 | 0 | 4 | 119.3067 | 14.6327 | 828.6344 |
| <i>Hebius parallelum</i> | Colubridae | NT | 0 | NA | 599.5 | 0.8057 | 0 | 3 | 2 | 0 | 3 | 39.0123 | 6.9991 | 517.9218 |
| <i>Hebius popei</i> | Colubridae | LC | 0 | LC | 512.5 | 1.0217 | 0 | 2 | 2 | 0 | 3 | 47.6291 | 19.9016 | 1525.8983 |
| <i>Hebius sauteri</i> | Colubridae | LC | 0 | LC | 467 | 1.026 | 0 | 3 | 2 | 0 | 4 | 110.8808 | 17.901 | 1437.2741 |
| <i>Hebius venningi</i> | Colubridae | VU | 0 | LC | 505 | 0.7719 | 0 | 1 | 4 | 0 | 2 | 4.4253 | 18.1914 | 1346.0622 |
| <i>Hebius vibakari</i> | Colubridae | VU | 1 | NA | 474 | 1.0699 | 0 | 1 | 4 | 0 | 4 | 80.84 | 3.5157 | 539.7987 |
| <i>Hemorrhois ravergieri</i> | Colubridae | VU | 1 | LC | 1027.5 | 1.0489 | 0 | 3 | 3 | 0 | 3 | 0.0525 | 0.6747 | 282.5 |
| <i>Herpetoreas platyceps</i> | Colubridae | VU | 1 | NA | 725 | 0.747 | 0 | 1 | 3 | 0 | 1 | 0.0123 | 4.0319 | 351.6535 |
| <i>Hydrophis caeruleascens</i> | Elapidae | NT | 0 | LC | 780 | 1.1081 | 1 | 4 | 1 | 1 | 1 | 0.1366 | 23.0833 | 1710.0588 |

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|--------------------------------|--------------|----|---|----|--------|--------|---|---|---|---|---|----------|---------|-----------|
| <i>Hydrophis curtus</i> | Elapidae | LC | 0 | NA | 905.5 | 1.1281 | 1 | 4 | 1 | 1 | 3 | 0.0761 | 23.0315 | 1567.1667 |
| <i>Hydrophis cyanocinctus</i> | Elapidae | NT | 0 | LC | 1677.5 | 0.9677 | 1 | 4 | 1 | 1 | 2 | 0.3819 | 21.1901 | 1577.4719 |
| <i>Hydrophis fasciatus</i> | Elapidae | NT | 0 | LC | 1098.5 | 0.9973 | 1 | 4 | 1 | 1 | 1 | 0.1366 | 23.0833 | 1710.0588 |
| <i>Hydrophis gracilis</i> | Elapidae | NT | 0 | LC | 970.5 | 1.2234 | 1 | 4 | 1 | 1 | 2 | 0.1417 | 22.6317 | 1463.1905 |
| <i>Hydrophis ornatus</i> | Elapidae | LC | 0 | LC | 1088 | 1.0000 | 1 | 4 | 1 | 1 | 1 | 0.3855 | 21.156 | 1577.4719 |
| <i>Hydrophis peronii</i> | Elapidae | LC | 0 | NA | 1164 | 1.0000 | 1 | 4 | 1 | 1 | 1 | 0.3353 | 21.607 | 1577.4719 |
| <i>Hydrophis platurus</i> | Elapidae | LC | 0 | LC | 547 | 1.1451 | 1 | 4 | 1 | 1 | 2 | 0.7176 | 17.6516 | 1287.5479 |
| <i>Hydrophis stokesii</i> | Elapidae | LC | 0 | NA | 1400 | 0.75 | 1 | 4 | 1 | 1 | 2 | 0.383 | 21.1804 | 1577.4719 |
| <i>Hydrophis viperinus</i> | Elapidae | LC | 0 | NA | 1026 | 1.1737 | 1 | 4 | 1 | 1 | 1 | 0.4028 | 20.9842 | 1566.9468 |
| <i>Hypsiscopus plumbea</i> | Homalopsidae | VU | 0 | NA | 485.5 | 0.7948 | 1 | 2 | 1 | 1 | 2 | 5.9622 | 20.4365 | 1465.5394 |
| <i>Indotyphlops lazelli</i> | Typhlopidae | CR | 1 | NA | 125.25 | 0.5854 | 0 | 1 | 3 | 0 | 1 | 0.0001 | 22 | 1912.8333 |
| <i>Laticauda colubrina</i> | Elapidae | LC | 0 | LC | 1127.5 | 0.7756 | 1 | 1 | 1 | 0 | 4 | 0.0249 | 20.3667 | 1121.8571 |
| <i>Laticauda laticaudata</i> | Elapidae | NT | 0 | LC | 655 | 2.1566 | 1 | 4 | 1 | 0 | 4 | 0.3855 | 21.156 | 1577.4719 |
| <i>Laticauda semifasciata</i> | Elapidae | NT | 0 | NT | 1077 | 1.2794 | 1 | 3 | 1 | 0 | 3 | 0.2108 | 20.6474 | 1529.7872 |
| <i>Lycodon aulicus</i> | Colubridae | NT | 0 | NA | 730 | 1.0857 | 0 | 1 | 3 | 0 | 1 | 121.2906 | 14.3153 | 785.6571 |
| <i>Lycodon fasciatus</i> | Colubridae | LC | 0 | NA | 796 | 1.5031 | 0 | 1 | 5 | 0 | 3 | 243.8667 | 13.609 | 810.2665 |
| <i>Lycodon flavozonatus</i> | Colubridae | LC | 0 | LC | 1091.5 | 1.0288 | 0 | 1 | 2 | 0 | 1 | 93.1889 | 18.7831 | 1556.8738 |
| <i>Lycodon futsingensis</i> | Colubridae | NT | 0 | LC | 890 | 0.8737 | 0 | 1 | 3 | 0 | 1 | 22.2893 | 20.8658 | 1646.2896 |
| <i>Lycodon gongshan</i> | Colubridae | NT | 0 | DD | 862.5 | 1.2638 | 0 | 1 | 3 | 0 | 2 | 0.1416 | 13.2858 | 1217.375 |
| <i>Lycodon liuchengchaoi</i> | Colubridae | LC | 0 | NA | 747 | 1.0000 | 0 | 1 | 3 | 0 | 1 | 101.78 | 12.2622 | 528.6803 |
| <i>Lycodon multizonatus</i> | Colubridae | NT | 0 | DD | 592.5 | 0.782 | 0 | 3 | 3 | 0 | 1 | 1.9449 | 13.1 | 608.0667 |
| <i>Lycodon rosozonatus</i> | Colubridae | EN | 1 | DD | 1018 | 1.0861 | 0 | 1 | 2 | 0 | 1 | 1.3007 | 17.8305 | 1910.5833 |
| <i>Lycodon rufozonatus</i> | Colubridae | LC | 0 | LC | 1320 | 0.913 | 0 | 1 | 3 | 0 | 2 | 309.8638 | 14.9627 | 1089.0703 |
| <i>Lycodon ruhstrati</i> | Colubridae | LC | 0 | LC | 1005.5 | 0.962 | 0 | 1 | 3 | 0 | 2 | 138.2584 | 17.5232 | 1383.0331 |
| <i>Lycodon septentrionalis</i> | Colubridae | LC | 0 | NA | 1290 | 1.0000 | 0 | 1 | 2 | 0 | 2 | 2.8675 | 16.6 | 1741.3333 |
| <i>Lycodon subcinctus</i> | Colubridae | LC | 0 | LC | 537.5 | 1.0594 | 0 | 1 | 3 | 0 | 2 | 108.1401 | 17.7402 | 1304.786 |
| <i>Macropisthodon rudis</i> | Colubridae | LC | 0 | LC | 1026.5 | 0.776 | 0 | 2 | 3 | 1 | 3 | 142.5182 | 17.2778 | 1341.2969 |
| <i>Myrrophis bennettii</i> | Homalopsidae | LC | 0 | NA | 567.5 | 0.9739 | 1 | 4 | 1 | 1 | 2 | 0.1279 | 23.0839 | 1793.9714 |

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|------------------------------------|--------------|----|---|----|--------|--------|---|---|---|---|---|----------|---------|-----------|
| <i>Myrrophis chinensis</i> | Homalopsidae | VU | 0 | NA | 780 | 0.8705 | 1 | 4 | 1 | 1 | 2 | 5.2107 | 20.6461 | 1552.9247 |
| <i>Naja atra</i> | Elapidae | VU | 0 | VU | 1325 | 1.0076 | 1 | 3 | 3 | 0 | 2 | 79.6067 | 18.9785 | 1494.8802 |
| <i>Naja kaouthia</i> | Elapidae | EN | 0 | LC | 781 | 1.7404 | 1 | 3 | 3 | 0 | 5 | 16.2032 | 15.5035 | 1120.8107 |
| <i>Natrix natrix</i> | Colubridae | VU | 1 | LC | 935 | 0.7 | 0 | 3 | 2 | 0 | 2 | 166 | 7.7374 | 114.3155 |
| <i>Natrix tessellata</i> | Colubridae | LC | 0 | LC | 770.5 | 0.9025 | 0 | 3 | 2 | 0 | 6 | 8.8329 | 8.24 | 91.9128 |
| <i>Oligodon albocinctus</i> | Colubridae | NT | 0 | NA | 710.5 | 0.9846 | 0 | 3 | 3 | 0 | 1 | 39.0123 | 6.9991 | 517.9218 |
| <i>Oligodon catenatus</i> | Colubridae | NT | 0 | NA | 517.5 | 1.0097 | 0 | 3 | 3 | 0 | 1 | 54.04 | 19.9521 | 1592.9691 |
| <i>Oligodon chinensis</i> | Colubridae | LC | 0 | LC | 714.5 | 0.9602 | 0 | 3 | 3 | 0 | 1 | 135.7278 | 17.9914 | 1391.9874 |
| <i>Oligodon cinereus</i> | Colubridae | LC | 0 | LC | 541.5 | 0.9835 | 0 | 3 | 3 | 0 | 2 | 95.5263 | 17.5398 | 1276.5589 |
| <i>Oligodon fasciolatus</i> | Colubridae | NT | 0 | NA | 805.5 | 0.8798 | 0 | 3 | 3 | 0 | 1 | 39 | 16.3175 | 1039.7943 |
| <i>Oligodon formosanus</i> | Colubridae | NT | 0 | LC | 843.5 | 1.2644 | 0 | 3 | 3 | 0 | 2 | 100.9937 | 18.7096 | 1494.391 |
| <i>Oligodon joynsoni</i> | Colubridae | VU | 1 | LC | 935 | 1.0000 | 0 | 3 | 3 | 0 | 1 | 0.2318 | 22.4 | 1260.4 |
| <i>Oligodon lacroixi</i> | Colubridae | NT | 0 | VU | 463 | 0.9829 | 0 | 3 | 4 | 0 | 1 | 0.0766 | 15.4261 | 760 |
| <i>Oligodon lungshenensis</i> | Colubridae | NT | 0 | NT | 619.5 | 1.0114 | 0 | 3 | 3 | 0 | 2 | 0.1271 | 17.4411 | 1296.75 |
| <i>Oligodon ornatus</i> | Colubridae | NT | 0 | LC | 563 | 1.5078 | 0 | 3 | 3 | 0 | 2 | 48.3762 | 17.5081 | 1505.0936 |
| <i>Oocatochus rufodorsatus</i> | Colubridae | LC | 0 | LC | 1007.5 | 1.0773 | 0 | 3 | 2 | 1 | 3 | 237.2396 | 1.0667 | 846.1206 |
| <i>Ophiophagus hannah</i> | Elapidae | EN | 0 | VU | 2480 | 1.4136 | 1 | 3 | 3 | 0 | 5 | 61.9982 | 11.0793 | 1398.4663 |
| <i>Opisthotropis andersonii</i> | Colubridae | NT | 0 | NT | 500 | 1.0000 | 0 | 1 | 2 | 0 | 2 | 0.082 | 22 | 1912.8333 |
| <i>Opisthotropis balteata</i> | Colubridae | LC | 0 | NA | 937.5 | 1.1429 | 0 | 3 | 2 | 0 | 1 | 0.7699 | 3.5 | 617.9 |
| <i>Opisthotropis cheni</i> | Colubridae | NT | 0 | LC | 513.5 | 0.7955 | 0 | 1 | 2 | 0 | 3 | 0.1552 | 19.1732 | 1601.8889 |
| <i>Opisthotropis guangxiensis</i> | Colubridae | NT | 0 | NT | 425.5 | 0.8703 | 0 | 1 | 2 | 0 | 2 | 0.4169 | 19.9166 | 1669.4 |
| <i>Opisthotropis jacobii</i> | Colubridae | NT | 0 | DD | 478 | 0.9875 | 0 | 1 | 2 | 0 | 2 | 3.7492 | 17.8863 | 968.896 |
| <i>Opisthotropis kuatunensis</i> | Colubridae | LC | 0 | LC | 645 | 0.9545 | 0 | 1 | 2 | 0 | 2 | 28.8055 | 18.3013 | 1585.3044 |
| <i>Opisthotropis lateralis</i> | Colubridae | LC | 0 | LC | 337 | 1.2925 | 0 | 1 | 2 | 0 | 4 | 12.751 | 21.4227 | 1685.8439 |
| <i>Opisthotropis latouchii</i> | Colubridae | LC | 0 | LC | 533.5 | 0.8492 | 0 | 1 | 2 | 0 | 2 | 102.9086 | 18.1845 | 1500.575 |
| <i>Opisthotropis maculosa</i> | Colubridae | NT | 0 | DD | 513.5 | 1.0256 | 0 | 1 | 2 | 0 | 2 | 0.4078 | 22 | 1808.6957 |
| <i>Opisthotropis maxwelli</i> | Colubridae | NT | 0 | DD | 525.5 | 0.5524 | 0 | 1 | 2 | 0 | 1 | 38.0656 | 20.073 | 1633.2731 |
| <i>Oreocryptophis porphyraceus</i> | Colubridae | LC | 0 | NA | 1000.5 | 1.2333 | 0 | 1 | 3 | 0 | 5 | 349.1689 | 11.8387 | 790.6747 |

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|--------------------------------------|---------------|----|---|----|--------|--------|---|---|---|---|---|----------|---------|-----------|
| <i>Orientocoluber spinalis</i> | Colubridae | LC | 0 | NA | 924 | 1.0465 | 0 | 3 | 3 | 0 | 4 | 527.2146 | 7.1197 | 347.5533 |
| <i>Orthriophis moellendorffi</i> | Colubridae | EN | 0 | NA | 1902.5 | 1.0261 | 0 | 4 | 3 | 0 | 3 | 41.64 | 20.3437 | 1583.5401 |
| <i>Orthriophis taeniurus</i> | Colubridae | EN | 0 | NA | 1240 | 6.5841 | 0 | 4 | 3 | 0 | 2 | 401.8769 | 11.543 | 763.5468 |
| <i>Ovophis makazayazaya</i> | Viperidae | NT | 0 | LC | 945 | 0.7182 | 1 | 1 | 3 | 0 | 5 | 121.1266 | 18.0129 | 1412.1496 |
| <i>Ovophis monticola</i> | Viperidae | NT | 0 | LC | 642.5 | 0.7603 | 1 | 1 | 3 | 0 | 5 | 171.179 | 8.4162 | 453.2709 |
| <i>Ovophis tonkinensis</i> | Viperidae | LC | 0 | LC | 585 | 0.5811 | 1 | 1 | 3 | 0 | 1 | 45.18 | 20.6345 | 1603.8628 |
| <i>Paratapinophis praemaxillaris</i> | Colubridae | NT | 0 | LC | 846 | 0.7265 | 0 | 1 | 2 | 0 | 2 | 1.6665 | 19.9924 | 1668.1758 |
| <i>Pareas boulengeri</i> | Pareatidae | LC | 0 | LC | 562 | 0.7984 | 0 | 1 | 3 | 0 | 2 | 79.9317 | 16.2723 | 1292.142 |
| <i>Pareas carinatus</i> | Pareatidae | NT | 0 | LC | 580 | 0.9016 | 0 | 1 | 3 | 0 | 2 | 6.8809 | 19.3308 | 1405.6909 |
| <i>Pareas chinensis</i> | Pareatidae | LC | 0 | NA | 469 | 0.7867 | 0 | 1 | 3 | 0 | 3 | 211.2267 | 15.8255 | 1036.6678 |
| <i>Pareas formosensis</i> | Pareatidae | NT | 0 | LC | 535 | 1.2292 | 0 | 1 | 3 | 0 | 3 | 3.1479 | 20.3468 | 1115.4731 |
| <i>Pareas hamptoni</i> | Pareatidae | NT | 0 | LC | 452.5 | 0.6306 | 0 | 1 | 3 | 0 | 1 | 27.8547 | 18.0498 | 1182.9887 |
| <i>Pareas margaritophorus</i> | Pareatidae | NT | 0 | LC | 430.5 | 0.9793 | 0 | 1 | 3 | 0 | 1 | 22.3366 | 21.4308 | 1601.7141 |
| <i>Pareas monticola</i> | Pareatidae | NT | 0 | NA | 536 | 0.5952 | 0 | 1 | 3 | 0 | 1 | 39.0123 | 6.9991 | 517.9218 |
| <i>Pareas stanleyi</i> | Pareatidae | LC | 0 | DD | 506 | 0.8267 | 0 | 1 | 3 | 0 | 2 | 1.777 | 17.473 | 1592.6522 |
| <i>Plagiopholis blakewayi</i> | Colubridae | LC | 0 | LC | 457.5 | 0.83 | 0 | 3 | 3 | 0 | 1 | 51.6769 | 15.8067 | 1036.9459 |
| <i>Plagiopholis nuchalis</i> | Colubridae | VU | 1 | LC | 473.5 | 0.8496 | 0 | 3 | 3 | 0 | 1 | 16.763 | 18.0453 | 1137.4509 |
| <i>Plagiopholis styani</i> | Colubridae | LC | 0 | LC | 362 | 1.0394 | 0 | 3 | 3 | 0 | 3 | 142.5068 | 17.2749 | 1348.4252 |
| <i>Protobothrops cornutus</i> | Viperidae | CR | 1 | NT | 1094 | 0.8387 | 1 | 4 | 3 | 0 | 1 | 98.2567 | 17.9089 | 1283.9456 |
| <i>Protobothrops jerdonii</i> | Viperidae | LC | 0 | LC | 1073 | 1.1721 | 1 | 4 | 3 | 1 | 3 | 66.5777 | 17.7221 | 1387.0108 |
| <i>Protobothrops mangshanensis</i> | Viperidae | CR | 1 | EN | 2232.5 | 0.8919 | 1 | 4 | 3 | 0 | 1 | 0.2086 | 18.3 | 1600 |
| <i>Protobothrops mucrosquamatus</i> | Viperidae | LC | 0 | LC | 1090 | 1.1206 | 1 | 1 | 3 | 0 | 4 | 253.3472 | 15.7864 | 1129.2401 |
| <i>Protobothrops xiangchengensis</i> | Viperidae | LC | 0 | LC | 1019.5 | 1.2936 | 1 | 1 | 4 | 0 | 2 | 7.4173 | 9.8234 | 738.8136 |
| <i>Psammodynastes pulverulentus</i> | Lamprophiidae | LC | 0 | NA | 539 | 0.9853 | 1 | 4 | 3 | 1 | 3 | 155.8089 | 11.8597 | 925.5718 |
| <i>Psammophis lineolatus</i> | Lamprophiidae | NT | 0 | NA | 922 | 1.0000 | 1 | 3 | 3 | 0 | 3 | 218.01 | 7.6065 | 146.57 |
| <i>Pseudoxenodon bambusicola</i> | Colubridae | LC | 0 | LC | 632 | 1.0586 | 0 | 3 | 3 | 0 | 2 | 57.0494 | 19.6336 | 1592.0427 |
| <i>Pseudoxenodon karlschmidti</i> | Colubridae | LC | 0 | LC | 595.5 | 1.5287 | 0 | 3 | 3 | 0 | 3 | 44.6369 | 19.3961 | 1459.8662 |
| <i>Pseudoxenodon macrops</i> | Colubridae | LC | 0 | LC | 1042 | 1.1618 | 0 | 3 | 3 | 0 | 1 | 204.5301 | 16.5129 | 1250.967 |

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|--------------------------------------|------------|----|---|----|--------|--------|---|---|---|---|---|----------|---------|-----------|
| <i>Pseudoxenodon stejnegeri</i> | Colubridae | LC | 0 | LC | 854 | 1.0957 | 0 | 3 | 3 | 0 | 1 | 106.3663 | 16.6089 | 1324.0659 |
| <i>Ptyas carinata</i> | Colubridae | EN | 0 | LC | 3011 | 1.0000 | 0 | 3 | 3 | 0 | 4 | 0.1893 | 19.6 | 1373 |
| <i>Ptyas dhumnades</i> | Colubridae | VU | 0 | NA | 2458 | 1.1505 | 0 | 3 | 3 | 0 | 2 | 376.9736 | 13.9893 | 889.7532 |
| <i>Ptyas korros</i> | Colubridae | VU | 0 | NA | 1638.5 | 1.2127 | 0 | 3 | 4 | 0 | 4 | 180.3599 | 17.7163 | 1358.4849 |
| <i>Ptyas mucosa</i> | Colubridae | EN | 0 | NA | 1972.5 | 1.0829 | 0 | 3 | 3 | 0 | 5 | 247.5622 | 12.1532 | 846.7202 |
| <i>Ptyas nigromarginata</i> | Colubridae | VU | 0 | NA | 1879 | 0.9272 | 0 | 3 | 3 | 0 | 2 | 105.8534 | 13.5089 | 612.1148 |
| <i>Python bivittatus</i> | Pythonidae | CR | 0 | VU | 7000 | 1.0000 | 0 | 1 | 2 | 0 | 5 | 77.0745 | 19.2341 | 1429.5394 |
| <i>Rhabdophis adleri</i> | Colubridae | NT | 0 | LC | 858.5 | 0.8522 | 0 | 3 | 3 | 0 | 4 | 1.6061 | 24.1583 | 1940.9545 |
| <i>Rhabdophis himalayanus</i> | Colubridae | VU | 0 | NA | 784.5 | 0.8634 | 0 | 3 | 3 | 0 | 2 | 1.0873 | 16 | 2529 |
| <i>Rhabdophis leonardi</i> | Colubridae | LC | 0 | LC | 599.5 | 1.0084 | 0 | 3 | 3 | 0 | 2 | 28.8717 | 13.3767 | 939.2007 |
| <i>Rhabdophis nigrocinctus</i> | Colubridae | NT | 0 | LC | 1080 | 1.1073 | 0 | 1 | 3 | 0 | 1 | 1.1101 | 19.8005 | 1781.7213 |
| <i>Rhabdophis nuchalis</i> | Colubridae | LC | 0 | LC | 659.5 | 0.8422 | 0 | 3 | 3 | 0 | 1 | 108.7808 | 12.2103 | 868.3665 |
| <i>Rhabdophis pentasupralabialis</i> | Colubridae | LC | 0 | NA | 622.5 | 0.9825 | 0 | 3 | 3 | 0 | 2 | 87.6 | 13.067 | 508.5282 |
| <i>Rhabdophis subminiatus</i> | Colubridae | LC | 0 | LC | 1054 | 0.8573 | 0 | 3 | 3 | 0 | 4 | 144.3007 | 17.2011 | 1313.5086 |
| <i>Rhabdophis swinhonis</i> | Colubridae | NT | 0 | LC | 547.5 | 0.8559 | 0 | 3 | 3 | 0 | 4 | 3.1479 | 20.3468 | 1115.4731 |
| <i>Rhabdophis tigrinus</i> | Colubridae | LC | 0 | NA | 1012.5 | 0.8276 | 1 | 3 | 3 | 0 | 4 | 654.1046 | 9.3533 | 634.7726 |
| <i>Sibynophis chinensis</i> | Colubridae | LC | 0 | LC | 766.5 | 1.1963 | 0 | 3 | 3 | 0 | 3 | 222.3571 | 16.6118 | 1244.3817 |
| <i>Sibynophis collaris</i> | Colubridae | LC | 0 | LC | 643.5 | 0.9266 | 0 | 3 | 3 | 0 | 2 | 37.3849 | 15.1897 | 995.4202 |
| <i>Sinomicrurus hatori</i> | Elapidae | VU | 0 | NA | 1000 | 1.0000 | 1 | 1 | 3 | 0 | 2 | 3.6193 | 20.347 | 1115.3512 |
| <i>Sinomicrurus kelloggi</i> | Elapidae | LC | 0 | LC | 637.5 | 0.9676 | 1 | 1 | 3 | 0 | 1 | 101.9733 | 18.5198 | 1435.0218 |
| <i>Sinomicrurus maccllelandi</i> | Elapidae | VU | 0 | NA | 670 | 0.8873 | 1 | 1 | 3 | 0 | 2 | 308.9689 | 11.5326 | 773.7471 |
| <i>Sinomicrurus sauteri</i> | Elapidae | VU | 0 | LC | 865 | 1.3067 | 1 | 1 | 3 | 0 | 2 | 1.7079 | 20.6253 | 1093.6667 |
| <i>Sinonatrix aequifasciata</i> | Colubridae | VU | 0 | LC | 1012.5 | 0.8493 | 0 | 3 | 2 | 0 | 3 | 119.4997 | 1.8055 | 1442.7153 |
| <i>Sinonatrix annularis</i> | Colubridae | VU | 0 | NA | 594 | 0.7731 | 0 | 3 | 2 | 1 | 2 | 202.1733 | 16.078 | 1091.2307 |
| <i>Sinonatrix percarinata</i> | Colubridae | VU | 0 | LC | 1158.5 | 0.7823 | 0 | 3 | 2 | 0 | 3 | 160.4574 | 1.7644 | 1355.0392 |
| <i>Sinonatrix yunnanensis</i> | Colubridae | VU | 0 | LC | 719.5 | 0.4609 | 0 | 3 | 2 | 0 | 4 | 5.9244 | 18.1038 | 1240.9003 |
| <i>Stichophanes ningshaanensis</i> | Colubridae | NT | 0 | DD | 637.5 | 0.9466 | 0 | 3 | 4 | 0 | 1 | 0.0576 | 12.0254 | 728.5 |
| <i>Thermophis baileyi</i> | Colubridae | CR | 0 | NT | 817.5 | 0.9123 | 0 | 3 | 2 | 1 | 2 | 4.9806 | 3.625 | 479.8794 |

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|------------------------------------|--------------|----|---|----|-------|--------|---|---|---|---|---|----------|---------|-----------|
| <i>Thermophis zhaoermii</i> | Colubridae | CR | 0 | EN | 820 | 0.9408 | 0 | 3 | 3 | 0 | 2 | 0.5332 | 7.1541 | 703.0357 |
| <i>Trachischium monticola</i> | Colubridae | VU | 1 | NA | 225 | 1.0000 | 0 | 1 | 3 | 0 | 1 | 0.6367 | 16 | 2358 |
| <i>Trachischium tenuiceps</i> | Colubridae | VU | 1 | NA | 300 | 1.0000 | 0 | 1 | 3 | 0 | 1 | 0.7699 | 3.5 | 617.9 |
| <i>Trimeresurus albolabris</i> | Viperidae | LC | 0 | LC | 936 | 0.6566 | 1 | 4 | 4 | 1 | 3 | 107.5954 | 17.8746 | 1327.8426 |
| <i>Trimeresurus gracilis</i> | Viperidae | NT | 0 | LC | 440 | 1.023 | 1 | 2 | 3 | 1 | 2 | 3.6193 | 20.347 | 1115.3512 |
| <i>Trimeresurus gumprechtii</i> | Viperidae | LC | 0 | LC | 1250 | 1.0833 | 1 | 1 | 5 | 1 | 1 | 39 | 16.3175 | 1039.7943 |
| <i>Trimeresurus sichuanensis</i> | Viperidae | LC | 0 | DD | 1000 | 0.8519 | 1 | 1 | 3 | 0 | 2 | 25.8567 | 15.7773 | 1111.2973 |
| <i>Trimeresurus stejnegeri</i> | Viperidae | LC | 0 | NA | 844 | 0.8388 | 1 | 1 | 4 | 1 | 2 | 347.206 | 14.2409 | 943.1363 |
| <i>Trimeresurus yunnanensis</i> | Viperidae | LC | 0 | NA | 855.5 | 0.7694 | 1 | 1 | 4 | 1 | 2 | 87.6 | 13.067 | 508.5282 |
| <i>Vipera berus</i> | Viperidae | EN | 0 | NA | 560 | 0.8983 | 1 | 2 | 3 | 1 | 3 | 184.74 | 7.3987 | 161.431 |
| <i>Vipera renardi</i> | Viperidae | EN | 0 | NA | 462.5 | 1.033 | 1 | 3 | 3 | 1 | 1 | 166 | 7.7374 | 114.3155 |
| <i>Xenochrophis flavipunctatus</i> | Colubridae | LC | 0 | LC | 1000 | 1.0000 | 0 | 3 | 2 | 0 | 4 | 146.4246 | 19.166 | 1546.5366 |
| <i>Xenochrophis piscator</i> | Colubridae | LC | 0 | NA | 897.5 | 0.8505 | 0 | 1 | 2 | 0 | 3 | 225.3522 | 12.34 | 940.9748 |
| <i>Xenopeltis hainanensis</i> | Xenopeltidae | NT | 0 | LC | 902.5 | 0.9202 | 0 | 1 | 3 | 0 | 2 | 38.6257 | 20.6493 | 1617.6111 |
| <i>Xenopeltis unicolor</i> | Xenopeltidae | VU | 0 | LC | 737.5 | 1.418 | 0 | 1 | 3 | 0 | 4 | 47.4866 | 18.4564 | 1289.4799 |

22 ^a Species and Family: based on China Biodiversity Red List 2015 and the Reptile Database (<http://www.reptile-database.org/>).

23 ^b China and IUCN: based on China Biodiversity Red List 2015 or IUCN Red List. Data Deficient (DD), null (NA), Least Concern (LC), Near Threatened
 24 (NT), Vulnerable (VU), Endangered (EN) and Critically Endangered (CR).

25 ^c RangeC: species was assessed under range-based criteria (1), or not (0).

26 ^d BR: dividing the maximum body length of male species by that of female species.

27 ^e Toxicity: venomous (1) or not (0).

28 ^f AP: nocturnal (0), crepuscular (1), diurnal (2), or nocturnal and diurnal (3).

29 ^g MH: aquatic (1), semi-aquatic (2), terrestrial (3), semi-arboreal (4) and arboreal (5).

30 ^h RM: oviparous (0) or viviparous (1).

31 ⁱ HS: the number of habitat types occupied by each specie.

32

33 **Table A2.** Hypotheses on the relationship between intrinsic and extrinsic factors and extinction risk.

| Factors | Prediction | Justification | References |
|---------------------|-------------------|--|--|
| Body length | + | Large body is related to low population densities, large home ranges, slow reproductive and recover rates | Reed & Shine, 2002; Murray et al. 2011; Böhm et al. 2016 |
| Body ratio | – | Females need comparatively more times than males to become mature, thus have lower reproduction rate | |
| Toxicity | + | Venomous snakes are prone to human persecution and thus may disappear from areas where human land use dominates. | Todd et al. 2017 |
| Activity period | + | Diurnal species are easier to be hunted by predators | Gittleman 1985; Fleagle 2013 |
| Microhabitat | – | Species that are highly aquatic and depend on streams and wetlands are more sensitive to human land use | Todd et al. 2017 |
| Reproductive mode | + | Viviparous species tend to be larger than oviparous species | Böhm et al. 2016; Todd et al. 2017 |
| Habitat specificity | – | Habitat specialists are at higher risk of extinction | Böhm et al. 2016 |
| Range size | – | Small-ranged species are vulnerable to demographic stochasticity, local catastrophes and inbreeding | Murray et al. 2011; Böhm et al. 2016 |
| Mean temperature | – | Ectotherms have slower life histories, and therefore lower reproduction in areas of lower temperatures | Böhm et al. 2016 |

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|--------------------|---|--|-----------------------|
| Mean precipitation | + | Areas with high levels of precipitation have higher productivity and potentially higher human disturbance | Böhm et al. 2016 |
| Minimum elevation | + | High minimum elevations suggest smaller, more restricted ranges | Böhm et al. 2016 |
| Exploitation index | + | Higher human exploitation index suggests potentially higher human disturbance and impact, such as for research, medicine, pet trade and food purpose | Ruland & Jeschke 2017 |

34

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47 **Table A3.** Main sources for assessing the values of minimum elevation (ME) and human exploitation index (EI). The year after IUCN in
 48 reference columns indicates the date when species is assessed as recorded in IUCN Red List (<http://www.iucnredlist.org/>). Human exploitation
 49 index: rarely or never exploited (0), occasionally exploited (1), commonly exploited (2) and frequently exploited (3).

| Species | ME | References | EI | References |
|---------------------------------|------|----------------------|----|--|
| <i>Achalinus ater</i> | 500 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Achalinus formosanus</i> | 1000 | Zhao 2006, IUCN 2017 | 0 | No report, IUCN 2017 |
| <i>Achalinus hainanus</i> | 750 | Zhao 2006, IUCN 2017 | 0 | No report, IUCN 2017 |
| <i>Achalinus jinggangensis</i> | 940 | Zhao 2006, IUCN 2014 | 0 | No report, IUCN 2014 |
| <i>Achalinus meiguensis</i> | 1200 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Achalinus niger</i> | 1000 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Achalinus rufescens</i> | 370 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Achalinus spinalis</i> | 300 | Zhao 2006, IUCN 2017 | 0 | No report, IUCN 2017 |
| <i>Ahaetulla prasina</i> | 197 | Zhao 2006, IUCN 2012 | 2 | Commonly collected as pet in China |
| <i>Amphiesma stolatum</i> | 215 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Amphiesmoides ornaticeps</i> | 150 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Archelaphe bella</i> | 1000 | Zhao 2006, IUCN 2018 | 1 | Pet occasionally, IUCN 2018 |
| <i>Atretium yunnanensis</i> | 800 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Azemioops feae</i> | 100 | Zhao 2006, IUCN 2012 | 2 | Frequently exploitation, Zhao 2006 |
| <i>Boiga cyanea</i> | 1780 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Boiga guangxiensis</i> | 200 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Boiga kraepelini</i> | 300 | Zhao 2006 | 1 | Pet small quantity, IUCN 2012 |
| <i>Boiga multomaculata</i> | 0 | Zhao 2006 | 1 | Pet occasionally, web report |
| <i>Bungarus fasciatus</i> | 20 | Zhao 2006, IUCN 2013 | 3 | Overharvesting, IUCN 2013, Zhou & Jiang 2005 |
| <i>Bungarus multicinctus</i> | 20 | Zhao 2006, IUCN 2012 | 3 | Overharvesting, IUCN 2012, Zhou & Jiang 2005 |
| <i>Calamaria pavementata</i> | 597 | Zhao 2006 | 0 | No report, IUCN 2012 |

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|----------------------------------|------|----------------------|---|---|
| <i>Calamaria septentrionalis</i> | 300 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Calamaria yunnanensis</i> | 1100 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Chrysopelea ornata</i> | 550 | Zhao 2006 | 0 | No report |
| <i>Coelognathus radiatus</i> | 300 | Zhao 2006, IUCN 2014 | 3 | Overharvesting, IUCN 2014, Zhou & Jiang 2005 |
| <i>Cyclophiops doriae</i> | 990 | Zhao 2006 | 0 | No report |
| <i>Cyclophiops major</i> | 200 | Zhao 2006, IUCN 2014 | 1 | Pet occasionally, IUCN 2014 |
| <i>Cyclophiops multicinctus</i> | 100 | Zhao 2006, IUCN 2012 | 1 | Pet occasionally, Personal communication |
| <i>Daboia siamensis</i> | 0 | Zhao 2006 | 3 | Overharvesting, IUCN 2012 |
| <i>Deinagkistrodon acutus</i> | 100 | Zhao 2006 | 3 | Overharvesting, Zhou & Jiang 2005 |
| <i>Dendrelaphis hollinrakei</i> | 100 | Zhao 2006 | 0 | No report, IUCN 2012 |
| <i>Dendrelaphis pictus</i> | 200 | Zhao 2006 | 1 | Pet occasionally, web report |
| <i>Elaphe anomala</i> | 460 | Zhao 2006 | 2 | Commonly exploited as food, Reports from the web |
| <i>Elaphe bimaculata</i> | 0 | Zhao 2006, IUCN 2012 | 1 | Pet small quantity, IUCN 2012 |
| <i>Elaphe carinata</i> | 100 | Zhao 2006 | 3 | Overharvesting, Zhou & Jiang 2005 |
| <i>Elaphe davidi</i> | 200 | Zhao 2006 | 1 | Food occasionally, Personal communication |
| <i>Elaphe dione</i> | 30 | Zhao 2006, IUCN 2017 | 1 | local wine, IUCN 2017 |
| <i>Elaphe schrenckii</i> | 0 | Zhao 2006 | 2 | Commonly exploited as food, Wang et al. 2011, The research status and conservation of the Snake, <i>Elaphe schrenckii</i> |
| <i>Elaphe zoigeensis</i> | 2800 | Zhao 2006, IUCN 2014 | 0 | No report, IUCN 2014 |
| <i>Emydocephalus ijimae</i> | -40 | Zhao 2006, IUCN 2010 | 1 | Medicine. Web reports |
| <i>Eryx miliaris</i> | 80 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Eryx tataricus</i> | 487 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Euprepiophis mandarinus</i> | 200 | Zhao 2006, IUCN 2012 | 3 | Overharvesting for food and pet, IUCN 2012 |
| <i>Euprepiophis perlacea</i> | 2000 | Zhao 2006, IUCN 2012 | 3 | Overharvesting, IUCN 2012 |
| <i>Gloydus brevicaudus</i> | 0 | Zhao 2006 | 3 | Over-collected, Zhou & Jiang 2005 |
| <i>Gloydus intermedius</i> | 620 | Zhao 2006 | 2 | Medicine, food, IUCN 2010 |
| <i>Gloydus lijianlii</i> | 0 | Zhao 2006 | 0 | No report |
| <i>Gloydus liupanensis</i> | 1950 | Zhao 2006 | 0 | No report |

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|-------------------------------|------|----------------------|---|---|
| <i>Gloydus monticola</i> | 3100 | Zhao 2006, IUCN 2012 | 0 | Unknown, IUCN 2012 |
| <i>Gloydus qinlingensis</i> | 1500 | Zhao 2006 | 0 | very occasionally exploited as pet |
| <i>Gloydus shedaensis</i> | 50 | Zhao 2006, IUCN 2012 | 0 | No report |
| <i>Gloydus strauchi</i> | 1500 | Zhao 2006 | 2 | Medicine, food. Web reports |
| <i>Gloydus ussuriensis</i> | 218 | Zhao 2006 | 2 | hunting as medicine and food, Web report |
| <i>Gonyosoma boulengeri</i> | 82 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Gonyosoma frenatus</i> | 200 | Zhao 2006 | 2 | Commonly exploited as pet, Reports from the web |
| <i>Gonyosoma prasinus</i> | 700 | Zhao 2006 | 2 | Commonly exploited as pet, Reports from the web |
| <i>Hebius atemporale</i> | 416 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Hebius bitaeniatum</i> | 800 | Zhao 2006 | 0 | No report, IUCN 2016 |
| <i>Hebius boulengeri</i> | 80 | Zhao 2006, IUCN 2016 | 0 | No report, IUCN 2016 |
| <i>Hebius craspedogaster</i> | 100 | Zhao 2006, IUCN 2016 | 0 | No report, IUCN 2016 |
| <i>Hebius johannis</i> | 1200 | Zhao 2006 | 0 | No report |
| <i>Hebius metusium</i> | 1200 | Zhao 2006, IUCN 2016 | 0 | No report, IUCN 2016 |
| <i>Hebius miyajimae</i> | 0 | Zhao 2006, IUCN 2016 | 0 | No report, IUCN 2016 |
| <i>Hebius modestum</i> | 920 | Zhao 2006 | 0 | No report, IUCN 2016 |
| <i>Hebius octolineatum</i> | 700 | Zhao 2006, IUCN 2016 | 1 | Pet occasionally, IUCN 2016 |
| <i>Hebius optatum</i> | 416 | Zhao 2006 | 0 | No report |
| <i>Hebius parallelum</i> | 1200 | Zhao 2006 | 0 | No report |
| <i>Hebius popei</i> | 281 | Zhao 2006, IUCN 2016 | 0 | No report, IUCN 2016 |
| <i>Hebius sauteri</i> | 680 | Zhao 2006, IUCN 2016 | 0 | No report, IUCN 2016 |
| <i>Hebius venningi</i> | 900 | Zhao 2006, IUCN 2016 | 0 | No report, IUCN 2016 |
| <i>Hebius vibakari</i> | 69 | Zhao 2006 | 0 | No report |
| <i>Hemorrhhois ravergeri</i> | 900 | Zhao 2006, IUCN 2017 | 0 | No report, IUCN 2017 |
| <i>Herpetoreas platyceps</i> | 2290 | Zhao 2006 | 0 | No report |
| <i>Hydrophis caeruleus</i> | -25 | Zhao 2006, IUCN 2010 | 1 | Medicine. Web reports |
| <i>Hydrophis curtus</i> | -30 | Zhao 2006 | 1 | Medicine. Web reports |
| <i>Hydrophis cyanocinctus</i> | -30 | Zhao 2006, IUCN 2010 | 1 | Medicine. Web reports |

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|--------------------------------|------|---|---|---|
| <i>Hydrophis fasciatus</i> | -12 | Zhao 2006, IUCN 2010 | 1 | Medicine. Web reports |
| <i>Hydrophis gracilis</i> | -30 | Zhao 2006, IUCN 2010 | 1 | Medicine. Web reports |
| <i>Hydrophis ornatus</i> | -22 | Zhao 2006, IUCN 2010 | 1 | Medicine. Web reports |
| <i>Hydrophis peronii</i> | -30 | Zhao 2006 | 1 | Medicine. Web reports |
| <i>Hydrophis platurus</i> | -10 | Zhao 2006, IUCN 2017 | 1 | Medicine. Web reports |
| <i>Hydrophis stokesii</i> | -30 | Zhao 2006 | 1 | Medicine. Web reports |
| <i>Hydrophis viperinus</i> | -30 | Zhao 2006 | 2 | Medicine. Web reports |
| <i>Hypsiscopus plumbea</i> | 20 | Zhao 2006 | 3 | Over-collected, Zhou & Jiang 2005 |
| <i>Indotyphlops lazelli</i> | 28 | Zhao 2006 | 0 | No report |
| <i>Laticauda colubrina</i> | -50 | Zhao 2006, IUCN 2010 | 0 | No report in China |
| <i>Laticauda laticaudata</i> | -80 | Zhao 2006, IUCN 2010 | 1 | Medicine. Web reports |
| <i>Laticauda semifasciata</i> | -20 | Zhao 2006, IUCN 2010 | 1 | Medicine. Web reports |
| <i>Lycodon aulicus</i> | 0 | Zhao 2006 | 0 | No report |
| <i>Lycodon fasciatus</i> | 900 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Lycodon flavozonatus</i> | 600 | Zhao 2006 | 1 | Pet occasionally, IUCN 2012 |
| <i>Lycodon futsingensis</i> | 200 | IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Lycodon gongshan</i> | 1517 | Chen et al. 2018, The Lycodon gongshan Found in Panzhihua City of Sichuan Province, China | 0 | No report, IUCN 2012 |
| <i>Lycodon liuchengchaoi</i> | 1230 | Zhao 2006 | 0 | No report |
| <i>Lycodon multizonatus</i> | 1000 | Zhao 2006, IUCN 2018 | 0 | No report, IUCN 2018 |
| <i>Lycodon rosozonatus</i> | 14 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Lycodon rufozonatus</i> | 0 | Zhao 2006, IUCN 2017 | 3 | Overharvesting, IUCN 2017, Zhou & Jiang 2005 |
| <i>Lycodon ruhstrati</i> | 600 | Zhao 2006, IUCN 2014 | 0 | No report, IUCN 2014 |
| <i>Lycodon septentrionalis</i> | 1500 | Zhao 2006 | 0 | No report |
| <i>Lycodon subcinctus</i> | 0 | Zhao 2006, IUCN 2012 | 1 | Medicine. Web reports |
| <i>Macropisthodon rudis</i> | 600 | Zhao 2006, IUCN 2012 | 3 | Trading massively, Zhou & Jiang 2005, IUCN 2012 |
| <i>Myrrophis bennettii</i> | 0 | Zhao 2006 | 0 | No report |
| <i>Myrrophis chinensis</i> | 20 | Zhao 2006 | 3 | Overharvesting, Zhou & Jiang 2005 |

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|------------------------------------|------|----------------------|---|--|
| <i>Naja atra</i> | 70 | Zhao 2006, IUCN 2014 | 3 | Overharvesting, IUCN 2014, Zhou & Jiang 2005 |
| <i>Naja kaouthia</i> | 450 | Zhao 2006, IUCN 2012 | 3 | Overharvesting, IUCN 2012 |
| <i>Natrix natrix</i> | 850 | Zhao 2006 | 0 | No report |
| <i>Natrix tessellata</i> | 90 | Zhao 2006, IUCN 2010 | 1 | Pet occasionally, IUCN 2010 |
| <i>Oligodon albocinctus</i> | 780 | Zhao 2006 | 0 | No report |
| <i>Oligodon catenatus</i> | 700 | Zhao 2006 | 0 | No report |
| <i>Oligodon chinensis</i> | 100 | IUCN 2012 | 1 | Pet occasionally, IUCN 2012 |
| <i>Oligodon cinereus</i> | 0 | Zhao 2006, IUCN 2010 | 1 | Pet occasionally, web report |
| <i>Oligodon fasciolatus</i> | 497 | Zhao 2006 | 0 | No report |
| <i>Oligodon formosanus</i> | 50 | Zhao 2006, IUCN2014 | 1 | Pet occasionally, Personal communication |
| <i>Oligodon joynsoni</i> | 300 | Zhao 2006, IUCN 2012 | 0 | Unknown, IUCN 2012 |
| <i>Oligodon lacroixi</i> | 1400 | Zhao 2006, IUCN 2012 | 1 | Pet occasionally, Personal communication |
| <i>Oligodon lungshenensis</i> | 900 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Oligodon ornatus</i> | 600 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Oocatochus rufodorsatus</i> | 0 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Ophiophagus hannah</i> | 225 | Zhao 2006, IUCN 2012 | 3 | Overharvesting, IUCN 2012, Zhou & Jiang 2005 |
| <i>Opisthotropis andersonii</i> | 300 | Zhao 2006, IUCN 2012 | 0 | IUCN 2012 |
| <i>Opisthotropis balteata</i> | 149 | Zhao 2006 | 1 | Pet occasionally, web report |
| <i>Opisthotropis cheni</i> | 480 | Zhao 2006 | 0 | No report, IUCN 2012 |
| <i>Opisthotropis guangxiensis</i> | 950 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Opisthotropis jacobi</i> | 500 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Opisthotropis kuatunensis</i> | 600 | Zhao 2006, IUCN 2014 | 0 | No report, IUCN 2014 |
| <i>Opisthotropis lateralis</i> | 100 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Opisthotropis latouchii</i> | 600 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Opisthotropis maculosa</i> | 190 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Opisthotropis maxwelli</i> | 425 | Zhao 2006 | 0 | No report, IUCN 2012 |
| <i>Oreocryptophis porphyraceus</i> | 200 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Orientocoluber spinalis</i> | 518 | Zhao 2006 | 1 | Mediation occasionally, Lei et al. 2006, A record of <i>Coluber spinalis</i> from Altai mountain |

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|--|------|----------------------|---|---|
| <i>Orthriophis moellendorffi</i> | 50 | Zhao 2006 | 3 | Overharvesting, Zhou & Jiang 2005 |
| <i>Orthriophis taeniurus</i> | 110 | Zhao 2006 | 3 | Overharvesting, Zhou & Jiang 2005 |
| <i>Ovophis makazayazaya</i> | 500 | Zhao 2006, IUCN 2012 | 1 | Pet occasionally, IUCN 2012 |
| <i>Ovophis monticola</i> | 315 | Zhao 2006 | 1 | Medicine |
| <i>Ovophis tonkinensis</i> | 800 | IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Paratapinophis praemaxillaris</i> | 475 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Pareas boulengeri</i> | 313 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Pareas carinatus</i> | 530 | Zhao 2006 | 0 | Unknown, IUCN 2012 |
| <i>Pareas chinensis</i> | 313 | Zhao 2006 | 0 | No report |
| <i>Pareas formosensis</i> | 313 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Pareas hamptoni</i> | 500 | Zhao 2006, IUCN 2012 | 1 | Pet occasionally, Personal communication |
| <i>Pareas margaritophorus</i> | 100 | Zhao 2006, IUCN 2012 | 1 | Pet occasionally, Personal communication |
| <i>Pareas monticola</i> | 1000 | Zhao 2006 | 0 | No report |
| <i>Pareas stanleyi</i> | 700 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Plagiopholis blakewayi</i> | 1300 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Plagiopholis nuchalis</i> | 1000 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Plagiopholis styani</i> | 1000 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Protobothrops cornutus</i> | 200 | IUCN 2012 | 1 | Occasionally, IUCN 2012 |
| <i>Protobothrops jerdonii</i> | 1350 | Zhao 2006, IUCN 2012 | 3 | Medicine, food, skin, Web reports |
| <i>Protobothrops mangshanensis</i> | 800 | Zhao 2006, IUCN 2012 | 2 | High demand for pet and well protected, IUCN 2012 |
| <i>Protobothrops mucrosquamatus</i> | 82 | Zhao 2006, IUCN 2010 | 3 | over-collected, Zhou & Jiang 2005 |
| <i>Protobothrops xiangchengensis</i> | 2750 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Psammodynastes pulverulentus</i> | 0 | Zhao 2006 | 0 | No report |
| <i>Psammophis lineolatus</i> | 900 | Zhao 2006 | 0 | No report |
| <i>Pseudoxenodon bambusicola</i> | 300 | Zhao 2006, IUCN 2012 | 1 | Pet occasionally, IUCN 2012 |
| <i>Pseudoxenodon karlschmidti</i> | 500 | Zhao 2006, IUCN 2014 | 0 | rarely collected as pet |

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|--------------------------------------|------|----------------------|---|---|
| <i>Pseudoxenodon macrops</i> | 500 | Zhao 2006, IUCN 2012 | 1 | Pet occasionally, Personal communication |
| <i>Pseudoxenodon stejnegeri</i> | 400 | Zhao 2006, IUCN 2014 | 1 | Pet occasionally, IUCN 2014 |
| <i>Ptyas carinata</i> | 500 | Zhao 2006, IUCN 2012 | 2 | Food, pet, IUCN 2012 |
| <i>Ptyas dhumnades</i> | 50 | Zhao 2006 | 3 | Overharvesting, Zhou & Jiang 2005 |
| <i>Ptyas korros</i> | 100 | Zhao 2006 | 3 | Food, medicine, over-collected, Zhou & Jiang 2005 |
| <i>Ptyas mucosa</i> | 150 | Zhao 2006 | 3 | Overharvesting, Zhou & Jiang 2005 |
| <i>Ptyas nigromarginata</i> | 1500 | Zhao 2006 | 2 | Food, pet, Personal communication |
| <i>Python bivittatus</i> | 10 | Zhao 2006, IUCN 2012 | 3 | Overharvesting, IUCN 2012 |
| <i>Rhabdophis adleri</i> | 82 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Rhabdophis himalayanus</i> | 900 | Zhao 2006 | 0 | No report |
| <i>Rhabdophis leonardi</i> | 1250 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Rhabdophis nigrocinctus</i> | 450 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Rhabdophis nuchalis</i> | 620 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Rhabdophis pentasupralabialis</i> | 1200 | Zhao 2006 | 0 | No report |
| <i>Rhabdophis subminiatus</i> | 850 | Zhao 2006, IUCN 2012 | 1 | Snake wine occasionally, IUCN 2012 |
| <i>Rhabdophis swinhonis</i> | 500 | Zhao 2006 | 0 | No report, IUCN 2012 |
| <i>Rhabdophis tigrinus</i> | 30 | Zhao 2006 | 1 | Pet occasionally, Personal communication |
| <i>Sibynophis chinensis</i> | 400 | Zhao 2006, IUCN 2014 | 1 | Pet occasionally, Personal communication |
| <i>Sibynophis collaris</i> | 830 | Zhao 2006 | 0 | No report, IUCN 2010 |
| <i>Sinomicrurus hatori</i> | 500 | Zhao 2006 | 0 | No report |
| <i>Sinomicrurus kelloggi</i> | 300 | Zhao 2006, IUCN 2012 | 0 | Unknown, IUCN 2012 |
| <i>Sinomicrurus macclellandi</i> | 215 | Zhao 2006 | 1 | Medicine. Web reports |
| <i>Sinomicrurus sauteri</i> | 500 | Zhao 2006, IUCN 2012 | 0 | IUCN 2012 |
| <i>Sinonatrix aequifasciata</i> | 100 | Zhao 2006, IUCN 2012 | 1 | Local food, IUCN 2012 |
| <i>Sinonatrix annularis</i> | 100 | Zhao 2006 | 3 | Over-collected, Zhou & Jiang 2005 |
| <i>Sinonatrix percarinata</i> | 100 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Sinonatrix yunnanensis</i> | 100 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |

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|------------------------------------|------|----------------------|---|---|
| <i>Stichophanes ningshaanensis</i> | 1400 | Zhao 2006, IUCN 2018 | 0 | No report, IUCN 2018 |
| <i>Thermophis baileyi</i> | 3000 | Zhao 2006, IUCN 2014 | 0 | No report, IUCN 2014 |
| <i>Thermophis zhaoermii</i> | 3362 | Zhao 2006, IUCN 2014 | 0 | No report, IUCN 2014 |
| <i>Trachischium monticola</i> | 600 | Zhao 2006 | 0 | No report |
| <i>Trachischium tenuiceps</i> | 800 | Zhao 2006 | 0 | No report |
| <i>Trimeresurus albolabris</i> | 281 | Zhao 2006, IUCN 2012 | 2 | Collected commonly, IUCN 2012 |
| <i>Trimeresurus gracilis</i> | 2000 | Zhao 2006, IUCN 2012 | 0 | Unknown, IUCN 2012 |
| <i>Trimeresurus gumprechtii</i> | 300 | IUCN 2012 | 1 | Occasionally Collected |
| <i>Trimeresurus sichuanensis</i> | 878 | Zhao 2006, IUCN 2012 | 0 | No report, IUCN 2012 |
| <i>Trimeresurus stejnegeri</i> | 150 | Zhao 2006 | 1 | Collected occasionally |
| <i>Trimeresurus yunnanensis</i> | 1400 | Zhao 2006 | 0 | No report |
| <i>Vipera berus</i> | 120 | Zhao 2006 | 1 | Occasionally collected |
| <i>Vipera renardi</i> | 600 | Zhao 2006 | 2 | Commonly collected as medicine. Kang 1980, Primary research on <i>Vipera renardi</i> |
| <i>Xenochrophis flavipunctatus</i> | 100 | Zhao 2006 | 2 | Commonly exploited as pet and food, IUCN 2012 |
| <i>Xenochrophis piscator</i> | 0 | Zhao 2006 | 1 | Medicine, local collected, He et al. 1999, A primary report on Guangzhou snake market |
| <i>Xenopeltis hainanensis</i> | 200 | Zhao 2006, IUCN 2012 | 1 | Occasionally collected as pet |
| <i>Xenopeltis unicolor</i> | 650 | Zhao 2006, IUCN 2012 | 3 | Overharvesting, IUCN 2012 |

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51 Main references

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54 Zhou, Z., & Jiang, Z. (2005). Identifying snake species threatened by economic exploitation and international trade in China. *Biodiversity and Conservation*, 14, 3525–3536.

56 **Table A4.** Spearman correlation matrices of intrinsic traits and extrinsic factors for all Chinese snakes, human exploited and unexploited species.
 57 The values in the upper right diagonal line represent the Spearman ρ , whilst the values in the lower left diagonal line represent the p values. The
 58 variables that were indicated as significant in univariate PGLS analyses were marked in bold. See Table A1 for variable abbreviations.

| (a) All snakes | BL | BR | Toxicity | AP | MH | RM | HS | RS | MT | MP | ME | EI |
|----------------|-----------|-------|----------|-----------|---------|-----------|---------|-----------|---------|---------|---------|-----------|
| BL | | 0.208 | 0.188 | 0.335 | 0.096 | -0.02 | 0.146 | 0.148 | -0.017 | 0.015 | -0.272 | 0.548 |
| BR | 0.003 | | 0.053 | 0.169 | 0.099 | -0.018 | 0.125 | 0.05 | 0.083 | 0.042 | -0.165 | 0.233 |
| Toxicity | 0.008 | 0.453 | | 0.158 | 0.016 | 0.563 | 0.079 | -0.024 | 0.085 | -0.023 | -0.259 | 0.277 |
| AP | < 0.001 | 0.017 | 0.025 | | -0.156 | 0.309 | 0.09 | 0.036 | -0.016 | -0.041 | -0.258 | 0.373 |
| MH | 0.176 | 0.163 | 0.818 | 0.027 | | -0.167 | -0.053 | 0.213 | -0.223 | -0.266 | 0.28 | 0.026 |
| RM | 0.782 | 0.798 | < 0.001 | < 0.001 | 0.018 | | -0.026 | -0.029 | 0.022 | -0.072 | -0.236 | 0.231 |
| HS | 0.039 | 0.077 | 0.263 | 0.202 | 0.455 | 0.71 | | 0.362 | -0.173 | -0.16 | -0.181 | 0.264 |
| RS | 0.036 | 0.481 | 0.732 | 0.609 | 0.002 | 0.683 | < 0.001 | | -0.439 | -0.391 | -0.098 | 0.408 |
| MT | 0.814 | 0.242 | 0.229 | 0.82 | 0.001 | 0.755 | 0.014 | < 0.001 | | 0.806 | -0.316 | -0.056 |
| MP | 0.831 | 0.556 | 0.746 | 0.567 | < 0.001 | 0.307 | 0.023 | < 0.001 | < 0.001 | | -0.277 | -0.053 |
| ME | < 0.001 | 0.019 | < 0.001 | < 0.001 | < 0.001 | 0.001 | 0.01 | 0.164 | < 0.001 | < 0.001 | | -0.426 |
| EI | < 0.001 | 0.001 | < 0.001 | < 0.001 | 0.712 | 0.001 | < 0.001 | < 0.001 | 0.43 | 0.454 | < 0.001 | |

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| (b) exploited snakes | BL | BR | Toxicity | AP | MH | RM | HS | RS | MT | MP | ME |
|----------------------|-----------|-------|----------|-----------|--------|-----------|--------|--------|-------|-------|--------|
| BL | | 0.167 | -0.078 | 0.243 | 0.16 | -0.274 | 0.085 | -0.036 | 0.128 | 0.189 | -0.035 |
| BR | 0.097 | | -0.085 | 0.136 | 0.000 | -0.158 | 0.068 | -0.041 | 0.068 | 0.089 | -0.065 |
| Toxicity | 0.44 | 0.402 | | 0.097 | -0.158 | 0.519 | -0.008 | -0.247 | 0.192 | 0.175 | -0.254 |

| | | | | | | | | | | | |
|-----------|-------|-------|---------|-------|---------|--------|--------|---------|---------|---------|--------|
| AP | 0.015 | 0.177 | 0.338 | | -0.304 | 0.231 | -0.097 | -0.247 | 0.23 | 0.196 | -0.241 |
| MH | 0.112 | 0.999 | 0.116 | 0.002 | | -0.316 | 0.056 | 0.388 | -0.26 | -0.318 | 0.509 |
| RM | 0.006 | 0.117 | < 0.001 | 0.021 | 0.001 | | -0.184 | -0.179 | 0.138 | 0.073 | -0.268 |
| HS | 0.4 | 0.504 | 0.938 | 0.337 | 0.581 | 0.067 | | 0.347 | -0.272 | -0.266 | 0.025 |
| RS | 0.724 | 0.683 | 0.013 | 0.013 | < 0.001 | 0.075 | | | -0.629 | -0.637 | 0.216 |
| MT | 0.204 | 0.502 | 0.056 | 0.021 | 0.009 | 0.171 | 0.006 | < 0.001 | | 0.885 | -0.357 |
| MP | 0.06 | 0.378 | 0.082 | 0.051 | 0.001 | 0.468 | 0.007 | < 0.001 | < 0.001 | | -0.356 |
| ME | 0.731 | 0.521 | 0.011 | 0.016 | < 0.001 | 0.007 | 0.805 | 0.031 | < 0.001 | < 0.001 | |

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| (c) unexploited snakes | BL | BR | Toxicity | AP | MH | RM | HS | RS | MT | MP | ME |
|------------------------|-------|-------|----------|-------|-----------|--------|--------|-----------|-----------|---------|--------|
| BL | | 0.029 | 0.223 | 0.165 | 0.084 | -0.003 | -0.032 | -0.104 | -0.151 | -0.175 | 0.059 |
| BR | 0.777 | | 0.094 | 0.073 | 0.21 | 0.068 | 0.131 | -0.002 | 0.118 | 0.014 | -0.091 |
| Toxicity | 0.025 | 0.348 | | 0.052 | 0.272 | 0.564 | 0.088 | 0.045 | -0.011 | -0.215 | -0.018 |
| AP | 0.098 | 0.467 | 0.607 | | -0.004 | 0.273 | 0.179 | 0.09 | -0.262 | -0.222 | 0.047 |
| MH | 0.401 | 0.035 | 0.006 | 0.968 | | 0.076 | -0.232 | -0.081 | -0.159 | -0.223 | 0.17 |
| RM | 0.974 | 0.501 | < 0.001 | 0.006 | 0.45 | | 0.092 | 0.014 | -0.137 | -0.237 | 0.049 |
| HS | 0.753 | 0.191 | 0.384 | 0.073 | 0.02 | 0.362 | | 0.194 | -0.046 | -0.036 | -0.148 |
| RS | 0.301 | 0.984 | 0.654 | 0.371 | 0.422 | 0.892 | 0.051 | | -0.262 | -0.212 | -0.029 |
| MT | 0.131 | 0.242 | 0.911 | 0.008 | 0.112 | 0.172 | 0.651 | 0.008 | | 0.738 | -0.475 |
| MP | 0.081 | 0.893 | 0.031 | 0.026 | 0.025 | 0.017 | 0.718 | 0.034 | < 0.001 | | -0.429 |
| ME | 0.558 | 0.365 | 0.856 | 0.642 | 0.089 | 0.626 | 0.138 | 0.772 | < 0.001 | < 0.001 | |

61 **Table A5.** The performance of PGLS models predicting the extinction risk of all Chinese snakes. The table shows model rank, change in AICc
62 from the top model (ΔAICc), model weight (w_i) and Adjusted- R^2 . See Table A1 for variable abbreviations. The two best models that included
63 and excluded geographic range size were highlighted in bold.

| | k | AICc | ΔAICc | w_i |
|-------------------------------|----------|---------------|---------------------|-----------------|
| BL + AP + EI + RM + RS | 6 | 552.4 | 0 | 0.4375 |
| AP + EI + RM + RS | 5 | 553.24 | 0.85 | 0.2866 |
| BL + EI + RM + RS | 5 | 554.56 | 2.16 | 0.1488 |
| EI + RM + RS | 4 | 555.61 | 3.22 | 0.0877 |
| BL + EI + RS | 4 | 559.02 | 6.62 | 0.016 |
| BL + AP + EI + RS | 5 | 559.21 | 6.81 | 0.0145 |
| EI + RS | 3 | 561.51 | 9.11 | 0.0046 |
| AP + EI + RS | 4 | 561.77 | 9.38 | 0.004 |
| BL + AP + RM + RS | 5 | 567.4 | 15 | 0.0002 |
| BL + AP + RS | 4 | 571.09 | 18.69 | 3.82E-05 |
| BL + RM + RS | 4 | 573.46 | 21.06 | 1.17E-05 |
| BL + RS | 3 | 574.33 | 21.93 | 7.57E-06 |
| AP + RM + RS | 4 | 577.93 | 25.53 | 1.25E-06 |
| BL + AP + EI + RM | 5 | 583.53 | 31.14 | 7.59E-08 |
| AP + RS | 3 | 583.69 | 31.29 | 7.03E-08 |
| AP + EI + RM | 4 | 585.12 | 32.72 | 3.44E-08 |

| | | | | |
|--------------|---|--------|-------|----------|
| BL + AP + RM | 4 | 585.54 | 33.14 | 2.79E-08 |
| RM + RS | 3 | 586.26 | 33.86 | 1.94E-08 |
| BL + EI + RM | 4 | 587.28 | 34.88 | 1.16E-08 |
| BL + AP + EI | 4 | 588.09 | 35.69 | 7.79E-09 |
| RS | 2 | 588.36 | 35.96 | 6.79E-09 |
| BL + AP | 3 | 589.08 | 36.69 | 4.73E-09 |
| EI + RM | 3 | 589.18 | 36.78 | 4.51E-09 |
| BL + EI | 3 | 589.21 | 36.81 | 4.45E-09 |
| AP + EI | 3 | 591.02 | 38.63 | 1.79E-09 |
| BL + RM | 3 | 591.41 | 39.01 | 1.48E-09 |
| EI | 2 | 592.04 | 39.65 | 1.08E-09 |
| BL | 2 | 592.05 | 39.66 | 1.07E-09 |
| AP + RM | 3 | 592.67 | 40.27 | 7.87E-10 |
| AP | 2 | 597.81 | 45.41 | 6.02E-11 |
| RM | 2 | 600.78 | 48.38 | 1.37E-11 |
| - | 1 | 602.29 | 49.89 | 6.42E-12 |

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68 **Table A6.** The interaction models between geographic range size and other important variables for all snakes and unexploited snakes in China.

| | Estimate | SE | <i>t</i> value | <i>p</i> |
|------------------------------------|----------|-------|----------------|----------|
| China Biodiversity Red List | | | | |
| <i>All snakes</i> | | | | |
| Range size × Body length | -0.014 | 0.004 | -3.764 | <0.001 |
| Range size × Reproductive mode | -0.080 | 0.041 | -1.950 | 0.053 |
| Range size × Activity period | -0.016 | 0.009 | -1.649 | 0.101 |
| Range size × Exploitation index | 0.030 | 0.016 | 1.918 | 0.546 |
| <i>Human unexploited snakes</i> | | | | |
| Range size × Microhabitat | -0.042 | 0.008 | -5.046 | <0.001 |
| Range size × Mean temperature | -0.064 | 0.008 | -8.312 | <0.001 |

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75 **Appendix 2. Patterns of extinction risk in Chinese snakes using the IUCN Red List criteria**

76 Among all Chinese snakes, only 132 species were assessed by the IUCN Red List (including 12 Data Deficient species), which are considerably
77 fewer than that in the China Species Red List (236 species). In addition, there are 55 endemic species assessed in the China Species Red List,
78 whereas the IUCN Red List only assessed the extinction status of 33 endemic Chinese snakes.

79 We found that the IUCN threat assessment of Chinese endemic snakes was significantly correlated with that of the China Species Red List
80 (Spearman $\rho = 0.28$, $p < 0.05$). However, there were large differences in extinction risk between these two Red List criteria. Among Chinese
81 snakes, 22 species were more endangered in China than in global. For example, *Thermophis baileyi* was Critically Endangered in China, but was
82 listed as Near threatened by the IUCN Red List (Table A1).

83 When using the criteria of IUCN Red List, range size, habitat specificity and body ratio were substantially important for Chinese endemic
84 snakes in the univariate PGLS analysis (Table A7). The best multivariable model for endemic snakes included all above variables (Table A8).
85 Range size ($w_+ = 0.99$, $p < 0.001$), habitat specificity ($w_+ = 0.95$, $p < 0.01$) and body ratio ($w_+ = 0.93$, $p < 0.01$) were also substantially important
86 in the confidence set (Table A9). When using the China Red List, however, high extinction risk in Chinese endemic snakes was associated with
87 diurnal activity period, habitat specialization, small range size, and high exploitation index (Table A9). The main reason for the differences in
88 predictors may result from the large difference in sample size: there are 55 endemic species assessed in the China Species Red List (MEP and
89 CAS 2015), whereas only 33 endemic species were assessed in the IUCN Red List.

91 **Table A7.** Results of univariate PGLS models predicting the extinction risk in Chinese endemic snakes using both Red List criteria. Significant
 92 correlations were marked in bold: ⁺ $p < 0.1$, * $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$. See variable abbreviations in Table A1.

| Variable | BL | BR | Toxicity | AP | MH | RM | HS | RS | MT | MP | ME | EI |
|----------------------------|------|---------------|----------|----------------|------|-------|----------------|-----------------|-------|-------|------|-------------------------|
| China Red List | | | | | | | | | | | | |
| Slope | 0.04 | -0.41 | 0.34 | 0.56 | 0.16 | 0.41 | -0.48 | -0.17 | -0.14 | 0.33 | 0.01 | 0.34 |
| <i>t</i> (<i>p</i>) | 0.11 | -0.44 | 0.90 | 3.87*** | 0.75 | 1.04 | -2.83** | -4.44*** | -0.44 | 0.90 | 0.76 | 1.83⁺ |
| IUCN Red List | | | | | | | | | | | | |
| Slope | 0.16 | -2.40 | 0.14 | 0.32 | 0.21 | -0.15 | -0.72 | -0.30 | -0.10 | -0.11 | 0.01 | 0.25 |
| <i>t</i> (<i>p</i>) | 0.27 | -1.96* | 0.23 | 1.66 | 0.63 | -0.21 | -3.55** | -4.73*** | -0.27 | -0.16 | 1.34 | 0.91 |

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100 **Table A8.** The performance of PGLS models predicting the extinction risk of Chinese endemic snakes using both Red List criteria. The table
 101 shows model rank, change in AICc from the top model ($\Delta AICc$), and model weight (w_i). The two best models that included and excluded range
 102 size were marked in bold.

| China Red List | k | AICc | $\Delta AICc$ | w_i | IUCN Red List | k | AICc | $\Delta AICc$ | w_i |
|--------------------------|----------|---------------|---------------|-----------------|---------------------|----------|--------------|---------------|-----------------|
| AP + EI + HS + RS | 5 | 147.05 | 0 | 0.4375 | HS + BR + RS | 4 | 86.16 | 0 | 0.8864 |
| AP + EI + RS | 4 | 148.42 | 1.38 | 0.22 | HS + RS | 3 | 91.75 | 5.59 | 0.0541 |
| AP + HS + RS | 4 | 149.04 | 2 | 0.1611 | BR + RS | 3 | 92.26 | 6.11 | 0.0419 |
| AP + RS | 3 | 150.13 | 3.09 | 0.0934 | RS | 2 | 94.84 | 8.69 | 0.0115 |
| EI + RS | 3 | 151.62 | 4.58 | 0.0444 | HS + BR | 3 | 96.25 | 10.1 | 5.70E-03 |
| EI + HS + RS | 4 | 151.82 | 4.77 | 0.0403 | HS | 2 | 101.47 | 15.31 | 4.00E-04 |
| HS + RS | 3 | 158.68 | 11.63 | 0.0013 | BR | 2 | 108.88 | 22.72 | 1.03E-05 |
| RS | 2 | 159.29 | 12.24 | 1.00E-03 | - | 1 | 110.48 | 24.32 | 4.64E-06 |
| AP + HS | 3 | 160.61 | 13.56 | 5.00E-04 | | | | | |
| AP + EI + HS | 4 | 161.09 | 14.04 | 4.00E-04 | | | | | |
| AP | 2 | 165.1 | 18.05 | 1.00E-04 | | | | | |
| AP + EI | 3 | 166.48 | 19.43 | 5.26E-05 | | | | | |
| EI + HS | 3 | 167.71 | 20.67 | 2.64E-05 | | | | | |
| HS | 2 | 168.96 | 21.92 | 1.42E-05 | | | | | |
| EI | 2 | 174.02 | 26.97 | 7.62E-06 | | | | | |
| - | 1 | 174.56 | 27.51 | 6.08E-07 | | | | | |

104 **Table A9.** Model-averaged parameter estimates (θ), unconditional standard errors (SE) and relative variable importance (w_+) for each variable in
 105 the 95% confidence set for endemic Chinese snakes using both Red List criteria.

| | w_+ | θ | SE | z value | p |
|------------------------------|-------|----------|-------|-----------|---------|
| <i>China Red List</i> | | | | | |
| (Intercept) | / | 1.164 | 0.596 | 1.952 | 0.051 |
| Range size | 1 | -0.152 | 0.037 | 4.078 | < 0.001 |
| Activity period | 0.91 | 0.363 | 0.128 | 2.836 | < 0.01 |
| Human exploitation | 0.74 | 0.417 | 0.164 | 2.543 | < 0.05 |
| Habitat specificity | 0.64 | -0.261 | 0.140 | 1.864 | 0.062 |
| <i>IUCN Red List</i> | | | | | |
| (Intercept) | / | 4.305 | 1.143 | 3.768 | < 0.001 |
| Range size | 0.99 | -0.221 | 0.061 | 3.598 | < 0.001 |
| Habitat specificity | 0.95 | -0.5 | 0.173 | 2.894 | < 0.01 |
| Body ratio | 0.93 | -2.385 | 0.847 | 2.815 | < 0.01 |

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108 **Appendix 3. Distribution of extinction risk among snake genera**

109 **Statistical analyses**

110 Following Bennett and Owens (1997), we applied a simulation test to determine whether extinction risk is randomly distributed among genera of
111 Chinese snakes. We could not perform such an analysis in family level because there are only eleven families in Chinese snakes based on the
112 classification of the Reptile Database (<http://www.reptile-database.org/>). We randomly picked 67 species from China species list because 67
113 species were classified as threatened by China Biodiversity Red List. We then calculated the proportion of species in each genus that had been
114 randomly picked. This procedure was iterated 10000 times, and the average number of genera in ten proportion classes across all simulation
115 replicates was taken as the predicted distribution of extinction risk (Bennett & Owens, 1997; Wang et al., 2018). If predicted proportional
116 distribution was significantly different from that of observed one, we would expect that extinction risk is non-randomly distributed among
117 genera. To test for such difference, we applied chi-square (χ^2) test on the sets of observed and predicted distribution datasets (Sokal & Rohlf,
118 1995; Bennett & Owens, 1997; Zar, 2010).

119 We then used the binomial distribution test to identify which families or genera contain unexpectedly large or small number of threatened
120 species (Bielby et al., 2006). The null hypothesis predicts that species in each family (or genus) are threatened randomly. Thereby, the
121 probability that a family (or genus) of N species contains K threatened species should follow the binomial distribution (Bennett & Owens, 1997),
122 where the overall proportion of species threatened across all genera is 0.3333 (67 threatened species out of 201 species). Because this question
123 was tested independently for each snake family (or genus), we thus calculated adjusted critical values using the Dunn–Sidak method (Sokal &

124 Rohlf, 1995). The conventional significance levels of 5% were correspondent to $p < 4.65 \times 10^{-3}$ for family and $p < 8.27 \times 10^{-4}$ for snake genus.

125 **Results and Discussion**

126 We found no significant difference between the observed and predicted frequency distributions of extinction risk among genera in Chinese
127 snakes (chi-square test, $\chi^2 = 2.82$, $df = 9$, $p = 0.971$). A main explanation for nonrandom extinction is that species in similar lineages may share
128 endangering traits, predisposing them to higher risk than others (Bennett & Owens 1997; Bielby et al. 2006; Jones et al. 2003). However, our
129 findings showed that extinction risk of snakes was randomly distributed across genera, which disagrees with previous research among taxa
130 (Bielby et al., 2006; Dulvy et al., 2014; Wang et al., 2018). Compared to these studies, we used a boarder range of species' biology traits and
131 external factors, and found that a group of factors, including small range size, large body size, oviparous reproduction, diurnal activity and high
132 human exploitation index, may have separate or combined effects on high extinction risk of Chinese snakes (Fig. 2). Therefore, species in
133 different lineages may have different extinction-promoting traits, which conversely predisposes them less statistically different than expected.

134 In addition, our study showed that none of the snake families or genera contained remarkably more or less threatened species than
135 expected by chance ($p > 0.05$; Tables A10, A11), despite that species in some families (e.g. Boidae) and genera (e.g. Ptyas) are all threatened.
136 This could be largely explained by the binomial distribution itself, which has relatively low statistics power when family is small (Sokal &
137 Rohlf, 1995; Bennett & Owens, 1997). As indicated by Bennett and Owens (1997), these small families may harbor some unique functional and
138 phylogenetic information. Thereby, any loss of species from these small families would lead to a disproportionately large loss of biodiversity
139 (Tonini et al., 2016).

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156 **Table A10.** Unusually highly threatened families within Chinese snakes. See Table 1 for more information.

| Family name | No. species in family | No. species threatened ^a | Proportion threatened | Binomial probability ^b |
|----------------|-----------------------|-------------------------------------|-----------------------|-----------------------------------|
| Boidae | 2 | 2 | 1.00 | 0.1111 |
| Colubridae | 123 | 41 | 0.33 | 0.0761 |
| Elapidae | 23 | 8 | 0.35 | 0.1707 |
| Homalopsidae | 3 | 2 | 0.67 | 0.2222 |
| Lamprophiidae | 2 | 0 | 0.00 | 0.4445 |
| Pareatidae | 8 | 0 | 0.00 | 0.0390 |
| Pythonidae | 1 | 1 | 1.00 | 0.3333 |
| Typhlopidae | 1 | 1 | 1.00 | 0.3333 |
| Viperidae | 28 | 9 | 0.32 | 0.1583 |
| Xenodermatidae | 8 | 2 | 0.25 | 0.2732 |
| Xenopeltidae | 2 | 1 | 0.50 | 0.4444 |

157 ^aData of 67 threatened species (Vulnerable, Endangered, and Critical Endangered) from the China Biodiversity Red List (MEP & CAS, 2015).

158 ^bBinomial probability $p = 0.3333$ (overall proportion of species threatened across all families).

159 * Adjusted significance using Dunn–Šidák correction at the 5% level ($p < 4.65 \times 10^{-3}$).

160 **Table A11.** Unusually highly threatened genera within Chinese snakes (top 13 abundant genera). See Table 1 and A10 for more information.

| Family name | No. species in genera | No. species threatened | Proportion threatened | Binomial probability |
|---------------|-----------------------|------------------------|-----------------------|----------------------|
| Hebius | 15 | 3 | 0.20 | 0.1299 |
| Lycodon | 12 | 1 | 0.08 | 0.0463 |
| Oligodon | 10 | 1 | 0.10 | 0.0867 |
| Hydrophis | 10 | 0 | 0.00 | 0.0174 |
| Opisthotropis | 10 | 0 | 0.00 | 0.0174 |
| Gloydus | 9 | 2 | 0.22 | 0.2341 |
| Rhabdophis | 9 | 1 | 0.11 | 0.1171 |
| Achalinus | 8 | 2 | 0.25 | 0.2732 |
| Pareas | 8 | 0 | 0.00 | 0.0390 |
| Elaphe | 7 | 4 | 0.57 | 0.1280 |
| Trimeresurus | 6 | 0 | 0.00 | 0.0878 |
| Ptyas | 5 | 5 | 1.00 | 0.0041 |
| Protobothrops | 5 | 2 | 0.40 | 0.3292 |

161 * Adjusted significance using Dunn–Šidák correction at the 5% level ($p < 8.27 \times 10^{-4}$).

162 **Appendix 4. Correlates of extinction risk in Chinese snakes when species classified under range-based criteria were excluded**

163 Small range size is likely to be the most important predictor for reptile species (Böhm et al., 2016), even in some cases when species listed as
164 threatened due to their small geographical range are excluded from the analyses (e.g. Cooper et al. 2008). In our study, 28 species were assessed
165 as threatened on the basis of their small range size (criteria B and D2). When excluding these species from the analyses, the traits of range size
166 ($w_+ = 1$), body length ($w_+ = 1$) and human exploitation index ($w_+ = 1$) were still important in determining the patterns of extinction risk (Table
167 A13, A14), which is consistent with our findings for all Chinese snakes (Fig. 2).

168 It is interesting that body ratio had no strong correlation with other traits (Table A12), but was substantially important in the 95%
169 confidence set when species classified under range-based criteria were excluded (Table A14). This indicates that when male body size of a given
170 adult snake is larger than that of female one, this species has considerably higher extinction risk. Therefore, our finding was contrary to the
171 previous expectation that females with a larger body size need comparatively more times than males to become mature, thus have lower
172 reproduction (Table A2). In general, large body size often means higher mortality under resource limitation (Rankin & Kokko 2007). Our study
173 already showed that species with larger body size were more likely to extinct due to the dual effects of habitat degradation and human
174 exploitation. With a larger body size, male snakes may put themselves at continuously high risk of human over-harvesting and resource
175 limitation before getting mature. This may conversely limit female access to a partner, and consequently decrease female mating rate and the
176 size of later generations (Bessa-Gomes et al. 2004), which overall increase their extinction probability. Moreover, small-bodied females often
177 have lower fecundity than larger ones (Goodwin et al. 2005), which may also play a role in the extinction vulnerability of snake population.

178 However, body ratio had a very low explanation on all Chinese species when including these species assessed due to small geographical
179 ranges (Table 2). This probably due to the artificial exclusion of species classified under range-based criteria, which changes the structure of
180 vectors and the relationship between body ratio and extinction risk. Moreover, it is also probably due to body ratio itself, which is correlated with
181 variables that are themselves positively and negatively related to extinction risk (as discussed above).

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195 **Table A12.** Spearman correlation matrix of intrinsic traits and extrinsic factors for all Chinese snakes. Species listed as threatened based only on
 196 their small range size have been omitted. The values in the upper right diagonal line represent the Spearman ρ , whilst the values in the lower left
 197 diagonal line represent the corresponding p values. See variable abbreviations in Table A1.

| | BL | BR | Toxicity | AP | MH | RM | HS | RS | MT | MP | ME | EI |
|-----------|-----------|-----------|----------|-----------|---------|--------|---------|-----------|---------|---------|--------|-----------|
| BL | | 0.167 | -0.078 | 0.243 | 0.160 | -0.274 | 0.085 | -0.036 | 0.128 | 0.189 | -0.035 | 0.351 |
| BR | 0.097 | | -0.085 | 0.136 | 0.000 | -0.158 | 0.068 | -0.041 | 0.068 | 0.089 | -0.065 | 0.149 |
| Toxicity | 0.44 | 0.402 | | 0.097 | -0.158 | 0.519 | -0.008 | -0.247 | 0.192 | 0.175 | -0.254 | 0.083 |
| AP | 0.015 | 0.177 | 0.338 | | -0.304 | 0.231 | -0.097 | -0.247 | 0.23 | 0.196 | -0.241 | 0.133 |
| MH | 0.112 | 0.999 | 0.116 | 0.002 | | -0.316 | 0.056 | 0.388 | -0.26 | -0.318 | 0.509 | 0.09 |
| RM | 0.006 | 0.117 | < 0.001 | 0.021 | 0.001 | | -0.184 | -0.179 | 0.138 | 0.073 | -0.268 | 0.03 |
| HS | 0.4 | 0.504 | 0.938 | 0.337 | 0.581 | 0.067 | | 0.347 | -0.272 | -0.266 | 0.025 | 0.286 |
| RS | 0.724 | 0.683 | 0.013 | 0.013 | < 0.001 | 0.075 | < 0.001 | | -0.629 | -0.637 | 0.216 | 0.155 |
| MT | 0.204 | 0.502 | 0.056 | 0.021 | 0.009 | 0.171 | 0.006 | < 0.001 | | 0.885 | -0.357 | -0.081 |
| MP | 0.06 | 0.378 | 0.082 | 0.051 | 0.001 | 0.468 | 0.007 | < 0.001 | < 0.001 | | -0.356 | -0.066 |
| ME | 0.731 | 0.521 | 0.011 | 0.016 | < 0.001 | 0.007 | 0.805 | 0.031 | < 0.001 | < 0.001 | | 0.081 |
| EI | < 0.001 | 0.139 | 0.413 | 0.187 | 0.371 | 0.768 | 0.004 | 0.124 | 0.426 | 0.515 | 0.425 | |

198 **Table A13.** Results of univariate PGLS models predicting extinction risk in Chinese snakes when species classified under range-sized criteria
 199 were excluded. Significant correlations were marked in bold: ⁺ $p < 0.1$, * $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$. See variable abbreviations in Table
 200 A1.

| Variable | BL | BR | Toxicity | AP | MH | RM | HS | RS | MT | MP | ME | EI |
|----------------------------|----------------|-------|----------|-------------------------|-------|-------|-------|---------------|------|------|------|----------------|
| Slope | 0.87 | 0.28 | -0.30 | 0.14 | -0.01 | -0.31 | -0.01 | -0.06 | 0.10 | 0.10 | 0.01 | 0.36 |
| <i>t</i> (<i>p</i>) | 4.63*** | 2.28* | -0.64 | 1.85⁺ | -0.03 | -1.26 | -0.05 | -2.22* | 0.86 | 0.78 | 0.11 | 5.34*** |

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211 **Table A14.** Model-averaged parameter estimates (θ), unconditional standard errors (SE) and relative variable importance (w_+) for each variable
 212 in the 95% confidence set for all Chinese snakes when species classified under range-sized criteria were excluded.

| Variables | w_+ | θ | SE | z value | p |
|--------------------------|-------|----------|-------|----------------|----------|
| intercept | / | -2.845 | 1.384 | 2.056 | < 0.05 |
| Range size | 1 | -0.102 | 0.027 | 3.743 | < 0.001 |
| Body length | 1 | 0.588 | 0.191 | 3.079 | < 0.01 |
| Human exploitation index | 1 | 0.328 | 0.073 | 4.466 | < 0.001 |
| Body ratio | 1 | 0.251 | 0.105 | 2.386 | < 0.05 |
| Activity period | 0.31 | 0.050 | 0.066 | 0.751 | 0.453 |